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**Modeling for sustainable groundwater management: interdependence and potential
complementarity of process-based, data-driven and system dynamics approaches**

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Abstract

Groundwater systems are vast natural water reservoirs used to support human water demands and ecosystem services. Various modeling approaches have been developed to help manage these complex highly-dynamic systems. This paper discusses the strengths and limitations of three modeling approaches, namely: process-based, data-driven and system dynamics modeling. For demonstration purposes, the three modeling approaches are applied to the Konya Closed Basin, a large agricultural region with semi-dry climate located in central Turkey. Process-based modeling is grounded in the theory-based representation of the governing processes but is somewhat limited by the computational effort and the difficulty of defining the required input parameters that characterize the heterogeneous aquifer system. Process-based models are shown to be powerful tools for resource management purposes provided climatic and water demand scenarios are accurately defined. Data-driven models are efficient tools for the management of groundwater resources but are highly dependent on the availability of large training data sets encompassing the spectrum of possible system responses. The high efficiency of surrogate modeling approaches makes them ideal tools for incorporation into applications such as real-time decision support systems and digital twin platforms. System dynamics modeling examines the groundwater exploitation problem within a socio-economic context that involves multiple stakeholders and their decision making. It combines groundwater flow models with socio-economics and endogenous decision rules to conduct scenario analysis and support policy development. The analyses and model demonstrations presented in this paper underscore the interconnectedness and complementarity of these three modeling approaches and the need for more integrated use of these modeling approaches for enhanced multi-sectoral management of groundwater systems.

Key words: process-based modeling, system dynamics modeling, data-driven modeling, water resources management, irrigated agriculture

44 **1. Introduction**

45 Groundwater is a widely accessible resource that is increasingly playing a vital role in providing freshwater
46 for domestic use, industry, and agriculture and for sustaining ecosystems (Ben Salem et al., 2023;
47 Berghuijs et al., 2022). Groundwater systems possess relatively large storage volumes, estimated to be as
48 much as 100 to 1000 times annual recharge, thus providing a high buffering capacity to mitigate against
49 seasonal as well as annual fluctuations in precipitation, particularly in arid and semi-arid regions (Margat
50 and van der Gun, 2013). As a result, global groundwater extractions have increased dramatically to more
51 than 1000 Gm³/year. This constitutes about 26% of the total global water use, including the water supply
52 for 40% of irrigated lands worldwide (Gleeson et al., 2016; Siebert et al., 2010). The utilization of
53 groundwater for irrigation has boosted economic growth and food production of many regions of the
54 world (Lejars et al., 2017, Butler and Whittemore, 2024). Nonetheless, the widespread overexploitation
55 of groundwater has also placed significant stress on this vital resource, leading to the depletion and
56 contamination of many aquifer systems worldwide. These stresses pose significant risks to human health,
57 socio-economic development and the environment, thereby threatening the sustainability of this vital
58 resource and raising issues of intergenerational inequity. Climate change is further intensifying these
59 challenges by instigating shifts in global temperatures and precipitation patterns which in turn affect
60 groundwater recharge (IPCC, 2021; Green et al., 2011; Secci et al., 2021; Secci et al., 2023). It is imperative
61 to recognize climate change as an external factor influencing groundwater quantity and quality and to
62 urgently assess the impact of future climate change scenarios on groundwater resources.

63 Groundwater systems are highly dynamic and complex systems with spatially variable properties, difficult
64 to estimate extraction and recharge rates, and highly temporal interactions with surface water bodies
65 (Saatsaz and Eslamian, 2020). Direct monitoring is often limited to small fractions of the aquifer volume.
66 As a result, modeling of subsurface systems has increasingly become an indispensable tool for the
67 management of groundwater resources. Modeling is utilized to address a wide range of issues relating to

68 both water quantity and quality, including the management of groundwater resources and assessing the
69 impact of climate change on this vital resource (e.g., Jódar et al., 2024; Mundetia et al., 2024; Rajaveni et
70 al., 2016).

71 The term modeling is very loosely used and can mean different things to different users (Konikow and
72 Bredehoeft, 1992). A model is defined as a representation of the real system and can be broadly classified
73 as physical, conceptual, mathematical or numerical. The development of models involves continuous
74 tradeoff between the incorporated level of realism, generalization and functionality (Levins, 1966). As a
75 result, all models involve some form of simplification and loss of precision and, hence, there is no single
76 correct model of a system (Cuddington et al., 2013). In the analysis of groundwater systems, like other
77 fields, widely varying modeling approaches exist. Physical models are typically simplified small-scale
78 representations of the real system normally constructed to examine specific flow or pollution transport
79 processes. Conceptual models depict the governing processes and relationships between variables.
80 Conceptual models are often utilized as an intermediate step for developing mathematical models that
81 provide a more quantitative representation of the flow and transport processes. Numerical models are
82 generally developed when model complexity prohibits the use of analytical tools for the solution of
83 mathematical models.

84 Modeling, the focus of the current study, varies widely in its purpose and underlying formulation. In the
85 context of groundwater resources management, three modeling approaches will be examined in this
86 study: process-based modeling (PBM), data-driven modeling (DDM) and system dynamics modeling
87 (SDM). PBM is built on fundamental concepts such as the principles of conservation of mass, energy or
88 momentum, often expressed in terms of governing partial differential equations. PBM relies on the latest
89 physical formulations for the relations between dependent and independent variables. Their main
90 challenges often include the computational effort involved and the need to quantitatively define the input
91 parameters describing the governing process.

92 In recent decades, with the advancement of computational power, there has been a growing reliance on
93 DDM to address complex problems. These models are particularly effective in capturing crucial
94 relationships between inputs, including geological properties, boundary conditions, pumping rates,
95 climatic data and outputs, such as groundwater levels, contaminant concentrations, or groundwater flow
96 rates. This capability enables swift simulations and facilitates decision support.

97 Whereas most groundwater models focus on the simulation of flow and contaminant transport, there are
98 other processes that also have significant impact on this resource. SDM (Forrester 1961, Sterman 2000),
99 simulates nonlinear interactions and constructs feedback systems that better explain the broader dynamic
100 nature of the systems including human, social and economic components. This modeling approach, which
101 has its roots in the management of complex industrial processes, has evolved into a scientific practice
102 with applications in various fields of research including socio-hydrological problems.

103 The rest of the paper is organized as follows. Section 2 gives an overview of the application of these three
104 modeling approaches (PBM, DDM and SDM) to groundwater related problems and how they evolved over
105 the years. Section 3 presents an application of these models that were developed for the management
106 of the groundwater system of the Konya Closed Basin, one of the major agricultural regions of Turkey that
107 is facing groundwater resources depletion due to overexploitation. Section 4 discusses the strengths,
108 limitations and complementarity of these three models

109 **2. Overview of Groundwater Modeling Approaches**

110 **2.1. Processed-Based Modeling**

111 PBM are mathematical representations of one or several processes or mechanisms that describe the
112 functioning of a well-delimited system (Buck-Sorlin, 2013). The processes are based on well-established
113 scientific understanding, typically expressed in terms of one or more differential equations. PBM can be
114 found in diverse disciplines such as forestry studies (Gonçalves et al., 2021), climate change modelling

115 (Wilson et al. 2021), soil sciences (Mu et al., 2020; Pierson et al., 2022), genetics and molecular biology
116 (Gross, 2024) and economics (Smart et al., 2007). PBM are also widely used in the field of hydrology with
117 the development of some models as well as debate among researchers on the challenges of these type of
118 models and how they can be overcome going back to more than 5 decades (Clark et al., 2017). In the
119 following subsection, we focus on the evolution of PBM relating specifically to the field of subsurface
120 hydrology.

121 The use of numerical models for the management of groundwater resources traces back to the middle of
122 the 20th century (Bredehoeft, 2012; de Marsily, 1986; Freeze, 2022). Earlier efforts focused on developing
123 analytical solutions of the governing equations for simplified input parameter distributions and boundary
124 conditions. With the development of computers, starting in the 1960's, numerical models that describe
125 groundwater flow became gradually available (Bredehoeft, 2012). In the 1970s, numerical techniques
126 such as the finite difference and finite elements were employed to solve the governing partial differential
127 equations. However, the progress in these endeavors was constrained by the computational capabilities
128 of the computers of that era. Further computational developments along with increased interest of
129 hydrogeologists in utilizing numerical models for the solution of groundwater flow and contaminant
130 transport problems led to the development of public domain PBMs such as MODFLOW (McDonald and
131 Harbaugh, 1988), SUTRA (Voss, 1984), and MT3D (Zheng, 1990). These publicly accessible models allow
132 for the simulation of the transient three-dimensional groundwater flow and contaminant transport for
133 complex geometries and boundary conditions and for spatially variable input parameters. Some of these
134 models remain in use to this day, albeit with a number of enhancements.

135 Models that can simulate saturated as well as unsaturated flow were also developed like the widely used
136 HYDRUS computer code (Šimůnek et al., 2022). Such models are particularly of interest to simulate aquifer
137 recharge and irrigation-related problems. Groundwater PBMs were also expanded to simulate multiphase
138 flows such as those related to the remediation of groundwater contaminated with non-aqueous phase

139 liquids (NAPLs). Computer codes capable of simulating multiphase flow include T2VOC (Pruess et al., 1993)
140 and UTCHEM (Delshad et al., 1996). The governing equations describing multiphase flow are highly non-
141 linear and coupled, including the occurrence of complex interphase mass transfer processes such as
142 evaporation and dissolution, making them particularly difficult to solve.

143 Recognizing that the input parameters required by these models are always associated by some
144 uncertainty due primarily to the heterogeneity of the subsurface environment and to the absence of
145 exhaustive data to fully define their spatial patterns, the field of stochastic subsurface hydrogeology
146 emerged with original works including Freeze (1975), Dagan (1989) and Gelhar (1993). Stochastic
147 hydrogeology rapidly evolved into a broad field of study with extensive research activity and field
148 applications (e.g., Kitanidis, 1997; Rubin, 2003) that are continuing to this very day (Renard et al., 2023).

149 PBM is also used for solving the inverse problem (de Marsily et al.1999, Carrera et al. 2005; Sun, 2014;
150 Zhou et al. 2014). In the typical forward groundwater flow problem, input parameters such as the
151 storativity and hydraulic conductivity are defined over the entire domain and the governing equations are
152 solved to yield the dependent variable such as the hydraulic head or concentration.

153 The tremendous development in the field of quantitative hydrogeology in recent decades has made PBM
154 an indispensable tool for the management of groundwater systems (Keller et al., 2023). PBM can help in
155 the interpreting of field data and for enhanced understanding of the complex spatially distributed
156 groundwater flow systems. Modeling has been used for the estimation of water budgets (e.g., Forootan
157 et al., 2014; Aderemi et al., 2022; Glose et al., 2022). The spatial distribution of budget components such
158 as the recharge may not be easily determined from measurements but can be estimated through model
159 calibrations. Models have also been used to estimate the safe and sustainable yields of pumping wells.
160 Such estimates can help in the sustainable use of groundwater resources and for extending aquifer
161 lifespans. Modeling has been shown to be useful as part of a Decision Support System that targets

162 improved management of limited water resources and/or protection of water resources from
163 contamination or seawater intrusion (Kazakis, 2018). Numerical models are also suitable for sensitivity
164 analyses and for risk assessment. With climate change expected to drastically influence water availability
165 in many regions of the world, groundwater modeling has been used to assess the impact of climate change
166 as well as other anthropogenic factors (Munir et al., 2024, Guevara-Ochoa et al., 2020).

167 Recognizing the importance of the interaction of surface and groundwater bodies, an increasing number
168 of coupled process-based surface/subsurface hydrology models have been developed (Shen and
169 Phanikumar, 2010; Ahmadi 2023). Numerous studies have shown that surface water bodies are strongly
170 interconnected with groundwater resources but that this connection is dynamic, bidirectional and
171 complex. Therefore, in many applications it is crucial to develop coupled surface-subsurface models for
172 enhanced management of water resources. The most commonly used models that can simulate surface
173 water/groundwater interactions are: SWAT-MODFLOW, GSFLOW, MIKE SHE, and HydroGeoSphere. Such
174 models can for example help in assessing the impact of land use and vegetation changes (Öztürk et al.,
175 2013), the impact of shallow water tables and irrigation practices on soil salinization (Nick et al., 2024)
176 and the benefits of different water management practices such as enhanced irrigation systems or
177 cropping changes (Wei and Bailey, 2019; Glose et al., 2022, Guevara-Ochoa et al., 2024).

178 **2.2. Data-Driven Modeling**

179

180 DDM is a particularly powerful type of surrogate modeling because it enables learning from data that is
181 too expensive to analyze using traditional methods, or, in the worst case, makes it possible to create
182 surrogate models for systems that would otherwise be controversial to model. Data-driven modeling
183 techniques use input-target datasets to map how the system works and to create a simplified model that
184 can be used to predict the system's behavior. These datasets can be generated from the complex system

185 itself through simulations or experiments, or they can be collected from historical data. Once the input-
186 target dataset is available, a data-driven modeling technique can be used to find linear or non-linear
187 relationships. This model can then be used to predict potential outputs for any given set of input
188 parameters.

189 The heterogeneous nature of aquifer properties introduces a substantial number of parameters which
190 can slow down the simulations of PBMs. This hinders their use in integrated modeling for real-time
191 decision support, especially for analyses involving multiple scenarios. A thorough analysis of surrogate
192 models in groundwater studies is presented by Asher et al. (2015) which sheds light on the prevalent use
193 of DDMs in predicting groundwater levels and addressing groundwater contamination issues.

194 Groundwater flow predictions leverage statistical methods, spatiotemporal geostatistics, and machine
195 learning techniques. Notably, these models require only an initial training phase that makes use of
196 historical or simulated data pertaining to the driving factors, such as precipitation and temperature, along
197 with their corresponding responses, such as groundwater levels. (Bierkens et al., 2001; Bloomfield and
198 Marchant, 2013; Varouchakis and Hristopulos, 2013; Secci et al., 2021; Secci et al., 2023)

199 Groundwater contamination represents another facet of DDMs applied to groundwater problems.
200 Specifically, DDMs have been considered in three key areas: contaminant transport modeling,
201 groundwater vulnerability assessment, and addressing inverse problems. The first area involves models
202 that aim to simulate and predict the movement of contaminants through aquifers. They integrate
203 extensive datasets, encompassing groundwater quality measurements, geological information,
204 contaminant source data, and vital hydrogeological parameters (Douglas and Efendiev, 2006; Opher et
205 al., 2009; He et al., 2019; Yu et al., 2020; Rodriguez-Galiano et al., 2014).

206 Groundwater vulnerability assesses the susceptibility of groundwater to contamination from diverse
207 sources. It involves understanding the hydrogeological characteristics, identifying potential contaminant

208 sources, and assessing pathways through which contaminants can enter the groundwater (Arthur et al.,
209 2007; Sajedi-Hosseini et al., 2018; Atenidegbe and Mogaji, 2023). For the solution of groundwater inverse
210 problems, Barati Moghaddam et al. (2021) and Gómez-Hernández and Xu (2022) have contributed critical
211 reviews that comprehensively analyze four decades of efforts made in this field.

212 **2.3. System Dynamics Modeling**

213 System dynamics emerged in the 1950s as a model-based discipline to explore dynamically complex
214 managerial problems. Since its inception, marked with the seminal works of Jay. W. Forrester (Forrester,
215 1961; Forrester, 1968), it has matured into a systems approach that addresses the root causes of chronic
216 socio-economic problems with an endogenous focus (Sterman, 2000; Barlas, 2009). Being a model-based
217 approach, advances in system dynamics have been highly influenced by developments in digital
218 computation, as well as in mathematics and statistics (Sterman, 2018). Today, applications of system
219 dynamics stretch over contemporary teams in corporate and business management, public policy and
220 resource management, and sustainability. System dynamics theory delves into the issues of systemic
221 model conceptualization/identification, formal model characterization, numeric computation, validation,
222 scenario building, policy analysis and implementation. Barlas (2009) comprehensively covers the
223 philosophical and methodological aspects of system dynamics. A short methodological outline for
224 modeling is presented in Feola et al. (2012).

225 In parallel to the theory and practice accumulated in system dynamics in the 1960s, Stanford Watershed
226 Model (Crawford and Linsley, 1966) and the Susquehanna River Basin Model (Hamilton, 1969) are
227 considered to be the first applications in water resources management. Winz et al. (2008) argues that the
228 systems approach to water resources management (Biswas, 1976; Rogers and Fiering, 1986) and
229 integrated water resources management have created a larger space for the uptake of SDM in water
230 resources management. Wurbs (1994), in their review of computer models for water resources

231 management, cite the stock and flow based STELLA II (High Performance Systems, 1985) simulation
232 platform of its time as one of the software that is in use by practitioners with the capabilities of “flexibility
233 in realistically representing real-world system concerns”, and “analysis, display and communication” but
234 also caution that the model building and maintenance can require large time and effort. Ford (1996),
235 Saysel et al. (2002), Stave (2003) and Martínez-Fernández and Selma (2004) can be cited among earlier
236 applications, sharing common principles and goals pertaining to system dynamics, such as dynamic
237 problem orientation, multi-sectoral model boundary selection comprising the hydrological as well as
238 socio-economic elements of a policy problem, feedback-rich model conceptualization, and model
239 construction, analysis and implementation with the purpose of learning and system design. Today, with
240 the emerging literature on food, water, energy nexus underlying the possible trade-offs and co-benefits
241 in resource provision (FAO, 2014; Simpson and Jewitt, 2019) and the socio-hydrological modeling (Blair
242 and Buytaert, 2016) demanding a holistic, multi-sectoral modeling approach for learning and
243 management, rather than “problem solving”, system dynamics is now facing larger opportunities to
244 respond to the needs as such in research and practice.

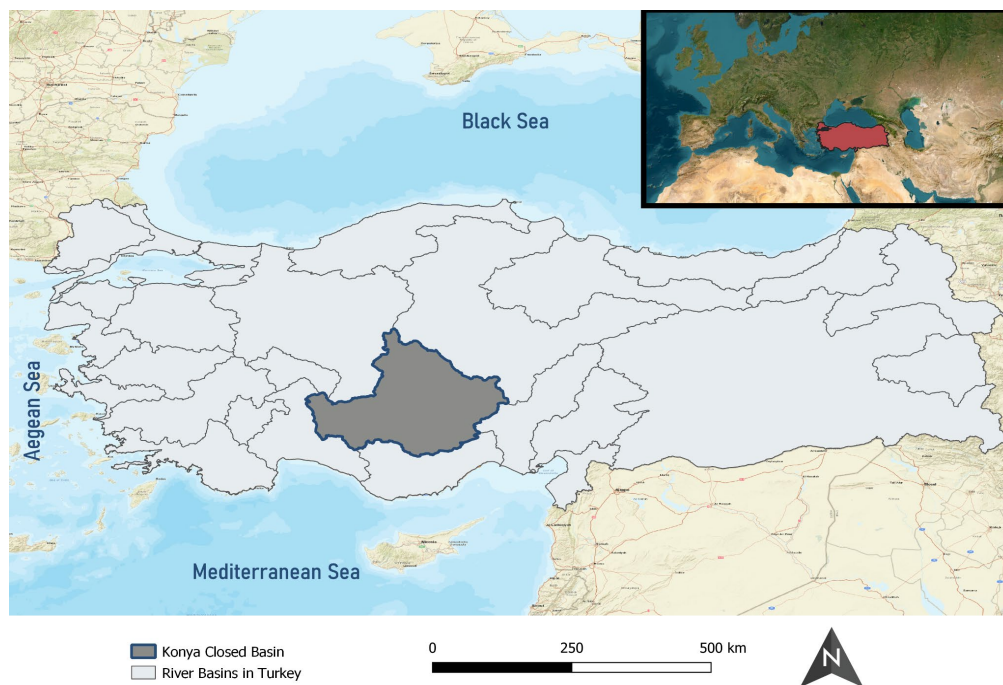
245 There are more recent inspiring applications of system dynamics in groundwater management studying
246 semi-arid, water stressed geographies in Iran, New Mexico, and Turkey. Bakhshianlamouki et al. (2020),
247 Ravar et al. (2020), and Naderi et al. (2021) adopt a nexus approach to water resources management by
248 developing dynamic feedback models for multi-sectoral policy analysis, the former with a focus on Lake
249 Urmia restoration and conservation, and the latter for sustainability measures in Gavkhuni and Qazvin
250 plains in Iran, respectively. Bai et al. (2021) study socio-hydrology in New Mexico with a focus on
251 agricultural water management in Rio Grande in New Mexico, analyzing policies extending from water
252 sector to agriculture. Langarudi et al. (2021) extend the work in Rio Grande to southern New Mexico.
253 Akhavan and Goncalves (2021) focus on groundwater use and sustainability in pistachio farming in the
254 Rafsanjan of Iran. Their analyses are particularly interesting to show “better before worse” consequences

255 of popularly received supply side solutions in water resources management. Mir et al. (2022) study the
256 “limits to growth” in the fossil water dependent agricultural sector in Sistan of Iran.

257 3. Application to the Konya Basin

258 3.1. Study Area and Data Availability

259 The three modeling approaches described above- PBM, DDM, and SDM were applied to the Konya Closed
260 Basin. The basin, located in Central Anatolia, is the biggest endorheic basin in Turkey, covering an area of
261 about 50,000 km² (Figure 1).



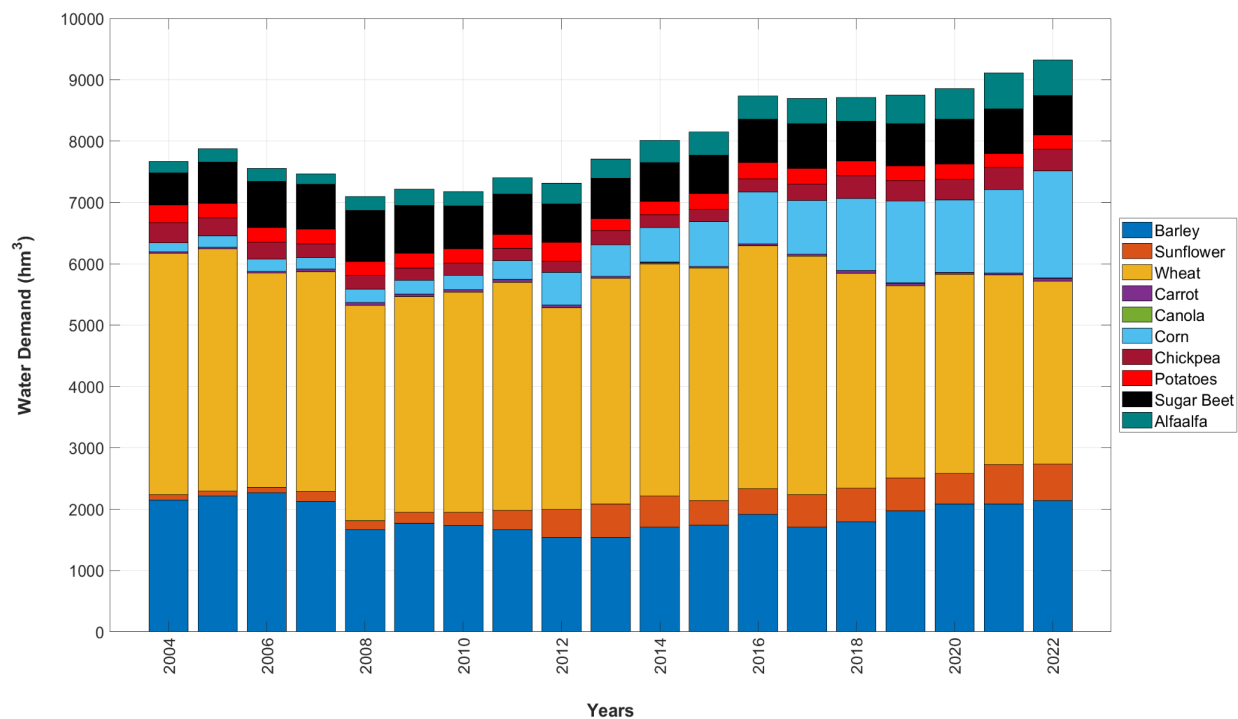
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Figure 1. Map of Konya Closed Basin

264 The basin is characterized by a semi-arid continental climate with hot and dry summers and cold and
265 moderately moist winters. The current average precipitation is about 380 mm/year with irregular
266 temporal and spatial distribution. The higher precipitation rates are over the mountainous areas located
267 in the southern and eastern portions of the basin with lower rates in the central plain. On the other hand,
268 the potential evapotranspiration over the basin is about 1400 mm/year (Bayari et al., 2009).

269 The basin is a major agricultural region with around 90% of the cultivated land used for grain production
 270 mainly wheat, barley, sugar beet and corn (Figure 2). Historically, mostly dry agriculture was practiced
 271 (Saygin et al., 2023). However, the past 2 decades witnessed a steady transition to more profitable and
 272 water-demanding crops such as corn and sugar beet (Yilmaz et al., 2021). Today, irrigation is the primary
 273 driver of water consumption in the basin, estimated to constitute more than 90% of the basin's water
 274 demands. Since surface water resources are limited, groundwater is widely used to supply the water
 275 deficiency. The number of wells in the basin is estimated as more than 100,000, with about two-thirds
 276 unlicensed (Yoloğlu, 2023). The overexploitation of the groundwater resource has led to a sharp decline
 277 in groundwater levels, placing significant pressure on the food production, economic viability, and
 278 ecosystem services of the basin (Özturk et al., 2023). Projected climate change is expected to further
 279 exacerbate the situation.



280

281

Figure 2. Water demand of the various crops cultivated in the Konya Closed Basin

282

283 The basin includes two primary aquifer systems: a shallow, fresh-water Neogene aquifer system and a
284 deeper saline aquifer system, separated by low permeability Paleogene rocks. The fresh-water Neogene
285 aquifer system has variable thickness extending to more than several hundred meters in some locations
286 (Bayari et al., 2009). Geologic maps and well logs show the presence of a thin alluvium, mostly silty loam,
287 overlaying the Neogene aquifer system with mountainous regions located along the edges of the basin.
288 The hydraulic conductivity in the basin varies between 0.43 and 593 m/day based on data from 81
289 pumping tests located mostly in the central valley of the basin. The current hydraulic head varies from
290 about 910 m above mean sea level in the central portions of the basin to more than 1200 m in the
291 mountainous region at the east boundary of the basin. Since pumping wells are generally unmetered,
292 groundwater extraction rates were estimated as described below indirectly from published official data
293 on the cultivated areas of each crop and from water demand data of different crops for the Konya region
294 (TUIK, 2023). These water demand calculations were performed on a monthly basis. Additionally, farmers'
295 decisions on crops, irrigation methods, and irrigation level based on multiple factors, such as costs, prices,
296 and groundwater and crop regulations, were also utilized in model development.

297 The three modeling approaches, which have different capabilities and hence use different datasets and
298 scenarios, are discussed in the following sections.

299

300 **3.2. Process-Based Model**

301 **3.2.1 Model Description**

302 The PBM was developed using the MODFLOW program coupled with the UZF1 module (Niswonger et al.,
303 2006) which simulates vertical water flow in the vadose zone along with horizontal flow in the underlying
304 aquifer system. The advantage of using the coupled saturated/unsaturated model is that it does not
305 require the definition of the spatially distributed recharge rate into the aquifer which is difficult to

306 estimate; instead, the model internally calculates the net recharge to the aquifer from the spatially and
307 temporally-dependent surface water (precipitation and irrigation) and the crop-dependent
308 evapotranspiration. Details of the model are given in (Yolođlu, 2023). The transmissivity and storativity
309 were estimated from a series of pumping tests and spatially interpolated over the entire basin using
310 kriging. As this is a closed basin, no-flow boundary conditions were assumed along the outer boundary.
311 Past daily precipitation distributions for the period from 2000-2023 were spatially interpolated from 18
312 meteorological stations located within the basin. As noted above, because pumping wells are generally
313 not metered, the groundwater extraction rates were estimated indirectly from crop cultivated areas and
314 the water demand data of the different crops (TUIK, 2023). Potential evapotranspiration rates were
315 defined based on published data for the Konya basin (TAGEM, 2017). Water demand calculations were
316 performed on a monthly basis.

317 The model was calibrated using monthly observed water level from 29 wells for the period from 2000-
318 2022. The irrigation efficiency, defined as ratio of the effective water use to the actual water withdrawal,
319 was designated as the primary calibration parameter. The irrigation efficiency is the product of
320 conveyance efficiency and field application efficiency (Brouwer et al., 1989). Typical values are in the range
321 of 0.2-0.3 for low efficiency conditions to about 0.6 for high efficiency irrigation systems (Rai et al., 2017).
322 The efficiency values determined from model calibration ranged from 0.3 for years 2000-2004 to 0.65 for
323 years 2015-2022. These values reflect the gradual transition by farmers to more efficient irrigation since
324 2010 due to water-authority incentives.

325 **3.2.2 Model Output and Analysis**

326 Figure 3a shows the simulated head distribution at the end of year 2022. It is observed that the heads in
327 the central area of the domain are lowest with flow from the boundary regions of the domain towards
328 the valley in the central portions of the domain. The average long-term drop in groundwater levels in the

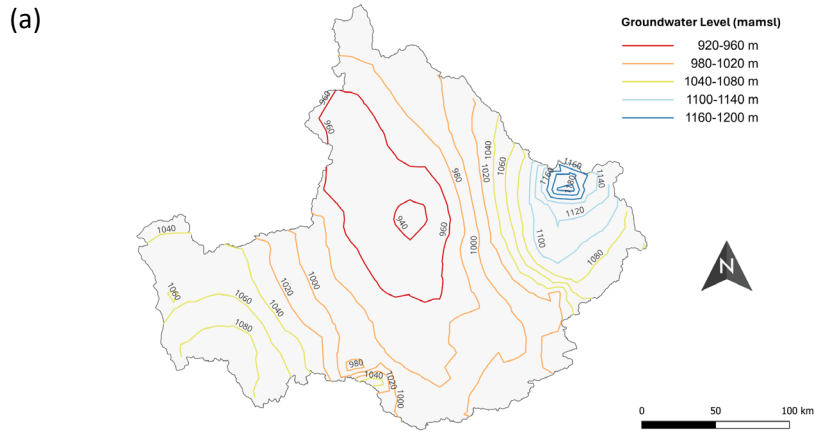
329 central area is estimated to be 0.62 m/yr. The significant decline in groundwater levels underlines the
330 unsustainable use of available water resources with potentially serious impacts on food production and
331 livelihood of the inhabitants of the region.

332 The calibrated PBM was then used to investigate the impacts of different groundwater use practices for
333 different climate change scenarios. Seventeen regional climate models from the EURO-CORDEX project
334 under two climate pathways (RCP 4.5 and RCP 8.5) were considered for the period of 2023-2040 (Todaro
335 et al., 2022). Five different water management scenarios were developed based on surveys with farmers,
336 NGOs, and water authorities. These management scenarios considered the impact of (i) enhanced
337 irrigation technology, (ii) roll back of corn cultivation, a recently expanded crop with relatively high water
338 demand, to 2005 levels and (iii) inter-basin water transfer. For each scenario, 34 simulations were
339 performed, each corresponding to one of the 17 available RCMs under the two climatic pathways, RCP
340 4.5 and 8.5. For demonstration purposes, results of two scenarios are presented here.

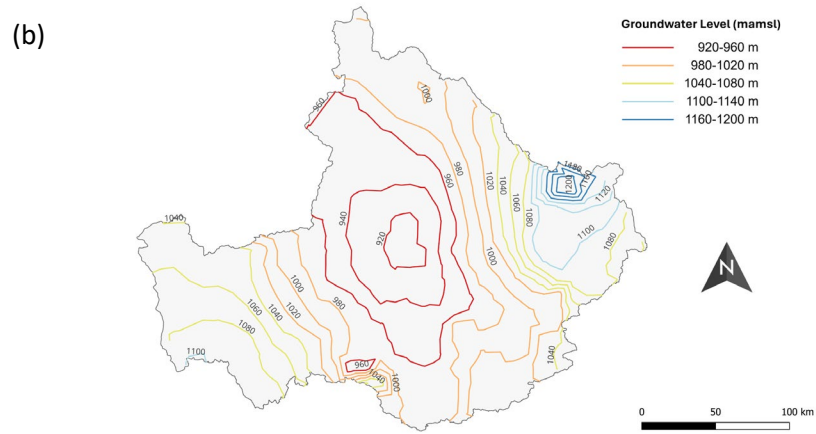
341 Figure 3b shows the water level distribution for December 2040 for the business as usual scenario,
342 characterized by the absence of substantial alterations in policy or behavioral pattern. This scenario
343 assumed that the trend in cropping that occurred in the years 2000-2022 continues into the future. The
344 irrigation efficiency for the business as usual scenario is also assumed to continue to increase at a rate
345 equal to that observed in the last two decades. The figure shows the simulated heads averaged over the
346 17 RCMs. The spatially averaged monthly head corresponding to the 17 RCMs is plotted in Figure 4a. The
347 blue dotted line on the figure shows the calibrated model result between 2000 and 2022 with the black
348 line depicting the median head for the RCP 4.5 and RCP 8.5 climatic pathways. A fan plot is used to depict
349 the variability of the simulation of the various RCMs under the two climatic pathways. The model predicts
350 that the rate of groundwater drop will gradually increase in the coming years as a result of expanded
351 irrigation lands and crop types. The average annual drop in groundwater level between the years 2023
352 and 2040 ranged between 0.02 and 0.39 m/year for individual climate projections, with an average of 0.28

353 m/year. Future rates of groundwater drop are smaller than historical values corresponding to years 2000-
354 2010 because of the switch to more efficient irrigation systems.

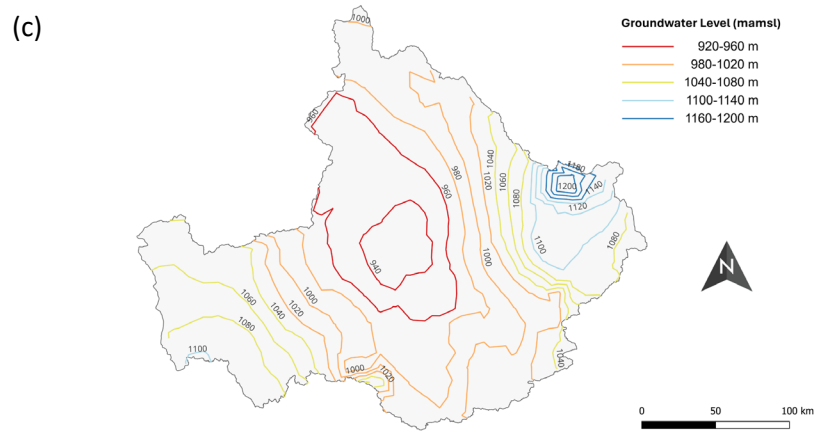
355 Figures 3c and 4b show the corresponding head distribution for the scenario corresponding to the roll
356 back of corn cultivation to 2005 levels and replacing it with wheat, a winter crop more suitable for the
357 semi-dry climate of Konya. This scenario also considers the continued trend in enhanced irrigation
358 efficiency that has been occurring in recent years. It is observed for this scenario that the average
359 groundwater levels are close to steady. These results demonstrate the type of analyses that can be
360 conducted with the PBM that can help decision makers and water authorities in managing available water
361 resources.



362



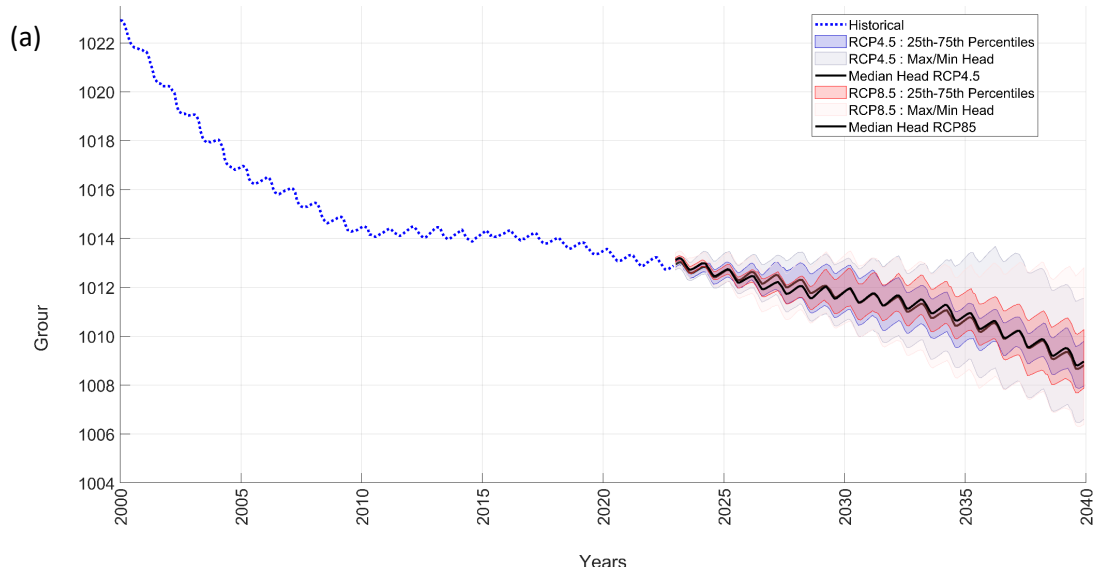
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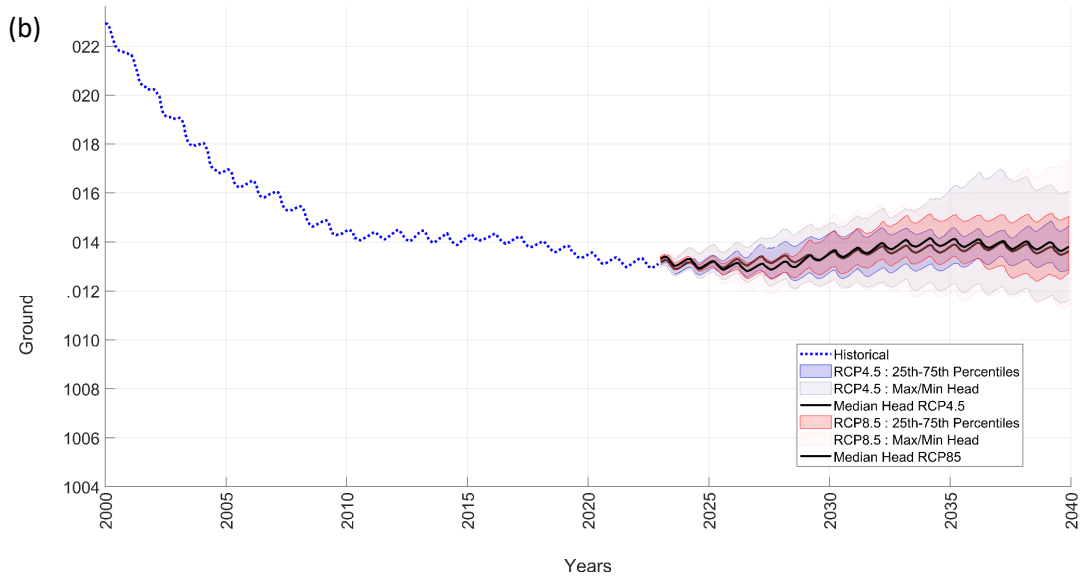
364

365 Figure 3. Simulated groundwater level in meters above mean sea level for (a) end of year 2022 at the end
 366 of the calibration process (b) end of year 2040 based on 17 RCMs under RCP4.5 climatic pathway for the
 367 business as usual scenario (c) end of year 2040 based on 17 RCMs under RCP4.5 climatic pathway for the
 368 reduced maize and enhanced irrigation scenario.

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370



371

372 Figure 4. Predicted groundwater level in meters above mean sea level averaged over the entire basin for

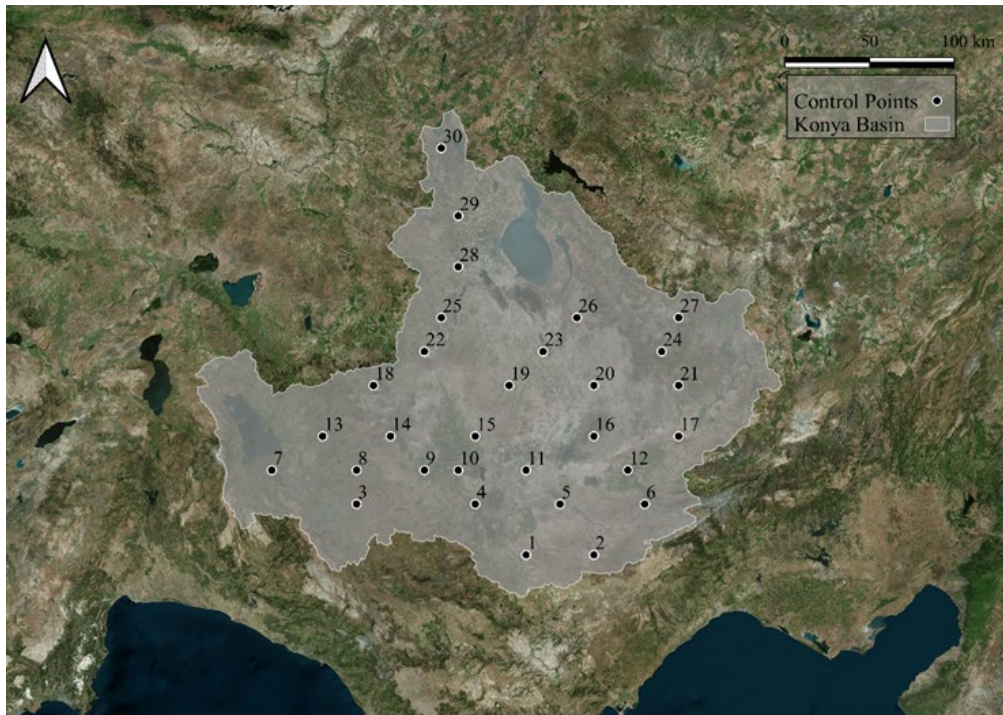
373 RCP 4.5 and 8.5 under (a) the business as usual (b) reduced maize and enhanced irrigation scenarios.

374

375 **3.3. Data-Driven Model**

376 **3.3.1. Model Description**

377 The implemented DDM , is an easy-to-use tool based on an artificial neural network(ANN). Details of the
378 mathematical concepts are available in Secci et al. (2022). The model predicts monthly groundwater levels
379 at 30 control points in the Konya Close Basin (Figure 5) until year 2040, considering various climate change
380 and agricultural practice scenarios. The climate scenarios consider changes in precipitation, while the
381 agricultural scenarios consider possible changes in crop types and irrigation patterns.



382
383 **Figure 5. Konya closed basin (Turkey) and the 30 control points**

384 In the application to Konya Closed Basin, the neural network consists of three layers: input, hidden and
385 output. The ANN has three input features, incorporating two multiplicative coefficients—one assigned to
386 historical precipitation (precipitation coefficient) and the other to historical crop water demand (crop
387 coefficient)—and time. A hidden layer comprising 10 neurons processes this input with the desired output
388 consisting of the groundwater levels at the 30 monitoring points (see Figure 5).

389 For training purposes, a dataset of 100 combinations of precipitation and crop coefficients was generated
390 using the Latin Hypercube Sampling method (McKay et al., 1979). The precipitation coefficient ranged
391 from 0.6 to 1.4, indicating a precipitation variation between -40% and +40% compared to the observed
392 period. Concurrently, the crop coefficient ranged from 0.75 to 1.25, corresponding to a water demand
393 varying between -25% and +25% of the historical one. For each combination, the PBM was executed from
394 2020-2040, generating monthly head data at the 30 monitoring points. The initial head distribution for
395 this simulation was based on the December 2019 head data obtained from the calibrated PBM.

396 The input data, comprising 100 combinations of precipitation, crop coefficients, and time, along with the
397 corresponding output (monthly heads at the 30 control points), constituted the dataset employed for
398 training the network. Since the simulation covers a period of 20 years (2020-2040), each combination
399 simulated by the full model produces 240 values of monthly heads for each of the 30 control points.
400 Therefore, for each control point, the total size of the training dataset is 24,000 (100 combinations
401 multiplied by 240 outputs each). The training dataset was partitioned into training (70%), validation (15%),
402 and testing (15%) sets and standardized for stabilizing neural network training, mitigating issues such as
403 vanishing gradients, and accelerating model convergence. The hyperparameters were calibrated manually
404 with a batch size of 128, number of epochs equal to 50. The early stopping was selected as regularization
405 criteria to prevent overfitting, while the learning and decay rates were set as 0.001 and 0.05, respectively.

406 **3.3.2. Model Output and Analysis**

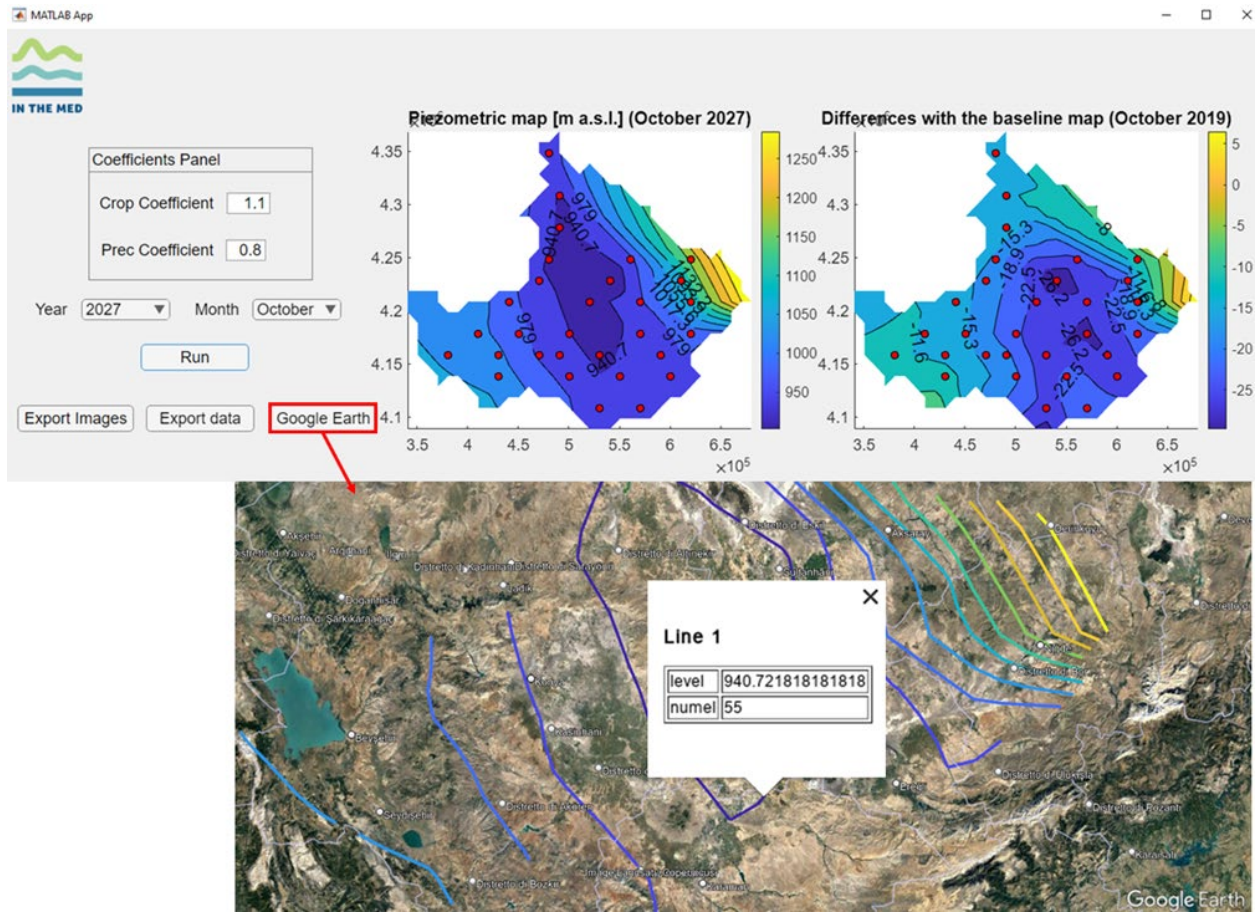
407 The mean square error of the overall network performance is 20.74 at the initial iteration, decreasing to
408 0.0061 at the final step of the training. This signifies a substantial decrease of at least three orders of
409 magnitude compared to the initial loss calculated with randomly initialized weights and biases. In addition,
410 the training, validation and test phases yielded means square error values of 0.0060, 0.0061 and 0.0062,
411 respectively. While these results might raise concerns about overfitting due to the strong performance in

412 the training phase, the consistent loss value associated with the validation dataset confirms effective
413 training. This is further supported by the neural network's robust ability to generalize when confronted
414 with data unseen during the testing phase.

415 Furthermore, the quality of the implemented model was also assessed in terms of the Coefficient of
416 Determination (R^2), which quantifies the proportion of the variance in the dependent variable that is
417 predictable from the independent variable(s). The R^2 values approach unity for the training, validation and
418 testing phases (0.9626, 0.9634, 0.9635) indicating a remarkably high level of predictability in capturing
419 the variance within the data.

420 The ultimate goal of the DDM was to develop an accessible and user-friendly tool that utilizes the trained
421 neural network to predict groundwater levels. Figure 6 shows the user interface of the ANN tool that was
422 developed in MATLAB (MathWorks, 2022). The left-hand panel allows users to simply select the
423 precipitation and crop coefficients and specify the period for which the results are required. Clicking the
424 "Run" button invokes the fully trained ANN, which uses the user-provided input features to generate the
425 desired output. The results are presented in the form of a piezometric head map. The monthly
426 groundwater levels at the 30 control points are provided by the ANN, while the final piezometric map is
427 generated by spatial interpolation of the network's output. Additionally, based on the user-defined
428 simulation period, the algorithm automatically creates a map that shows the differences between the
429 obtained piezometric map and the baseline map from 2019 for the selected month.

430 The tool allows the user to export the images and the data associated with the 30 control points and the
431 interpolated values to a text file. Finally, the resulting maps can be visualized as a contour plot
432 superimposed over Google Earth Pro (Google LLC, 2023).



433

434

Figure 6. Piezometric map generated by the ANN tool

435 3.4 The System Dynamics Model

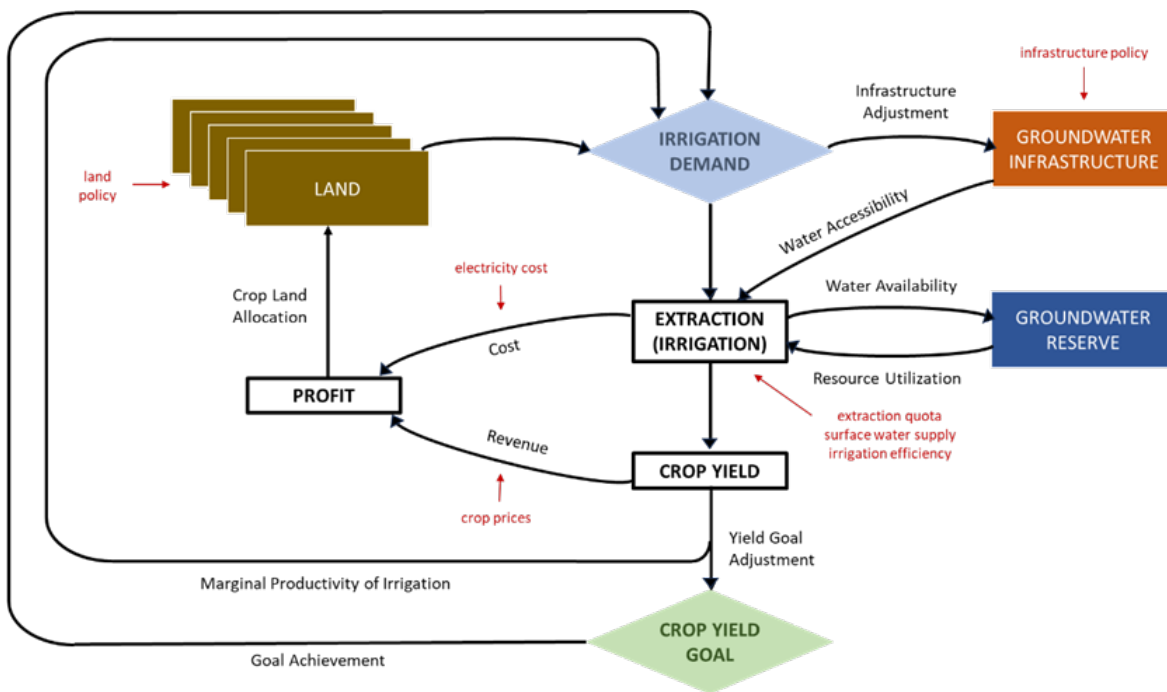
436 3.4.1 Model Description

437 The SDM explores the socio-economic and behavioral drivers of unsustainable groundwater use and aims
 438 to build a shared understanding of sustainable future pathways in the Çumra district of the Konya Closed
 439 Basin. Çumra is about 2200 km² and is considered one of the most agriculturally productive regions of the
 440 basin. To that end, the model was developed in parallel to three model group building workshops, in a
 441 living lab context (Andersen et al., 2007; Ceseracciu, 2023) with the participation of multiple stakeholders,
 442 including local farmers, water management authorities, chambers of agriculture, regional development
 443 administration, irrigation unions and cooperatives, academics, and environmental and agricultural NGOs.

444 First, desktop research was conducted for problem and stakeholder identification. Then, two field trips
445 were organized to acquaint the research team with the stakeholders and initiate the living lab, followed
446 by two participatory modeling workshops. The first workshop aimed to arrive at a shared understanding
447 of the problems and to identify potential policy intervention points, representing the ideas and
448 preferences of the diverse stakeholder groups. The second workshop focused on developing the seeds of
449 a conceptual model through a joint examination on the nature of the relationships between variables.
450 Building on the outputs of the workshops, the research team focused on formal model building, and then
451 organized a third field trip for testing and validation of the model with selected stakeholders. In a third
452 and final workshop, the stakeholders had the opportunity to explore the long-term impacts of multiple
453 policy alternatives (as specified in the previous workshops) on a simulation platform through a model
454 interface designed for that purpose and discuss the outcomes, reflecting back on their previous opinions.
455 Further details regarding the group model building process are presented in Uygur (2023).

456 Figure 7 shows the conceptual model, where the causal interactions between the variables are presented
457 with black arrows. Irrigation demand is driven by the crop land cover, farmers' crop yield goal, and
458 marginal productivity of irrigation, i.e., crop yield return to water. Whether the demanded irrigation water
459 can be supplied from groundwater depends on groundwater availability (remaining groundwater
460 reserves) and accessibility (capacity of existing extraction infrastructure). Groundwater component is
461 incorporated in the model with a single stock, representing the saturated zone. It is depleted with
462 extraction and fed by recharge and lateral flow. Groundwater extraction infrastructure includes wells and
463 pumps, which are adjusted annually based on irrigation demand. The crop yields are determined by the
464 level of irrigation. Farmers update (i) land allocation of crops based on the relative profitability of crops
465 and (ii) crop yield goals based on past experience. Production costs and generated revenues for the unit
466 production of each crop are compared to determine the profitability of crops relative to one another.
467 Though the total agricultural land is constant in the model, the size of the cultivated land changes due to

468 the fallow practice in non-irrigated lands, and the share of each crop is adjusted continuously; more land
 469 is spared for more profitable crops. The red arrows represent potential policy interventions, aiming either
 470 to control groundwater supply directly (for example, enforcing a groundwater extraction quota or
 471 restricting well digging), or to decrease the demand for groundwater (for example, incentivizing less water
 472 demanding crops or enforcing crop rotation schemes). A publicly available user-friendly interface is
 473 published on the web ([https://exchange.iseesystems.com/public/izel/inthemed-final-model---](https://exchange.iseesystems.com/public/izel/inthemed-final-model---eng/index.html#page1)
 474 [eng/index.html#page1](https://exchange.iseesystems.com/public/izel/inthemed-final-model---eng/index.html#page1)), on which stakeholders run their own policy experiments with various options
 475 such as crop rotation, surface water transfer, extraction limits, and repricing of crops.



476

477

Figure 7: System dynamics conceptual model

478 The model was developed on STELLA Architect (ISEE Systems, 2023). It comprises groundwater hydrology,
 479 as well as farmer decisions on crops, irrigation method, and level of irrigation based on multiple driving
 480 factors, such as costs, prices and groundwater and crop regulations. The time unit of the model is one
 481 year, and the simulation horizon is 36 years from 2004 to 2040. The presented outputs in Figure 8 are

482 simulated by solving the system of coupled ordinary differential equations using Euler's integration
483 method with a computation step equal to 1/32. The integration method and time steps were based on
484 criteria presented in Sterman (2000, Appendix A).

485 Prior to model based policy analysis, various validation tests were applied to the model. Socio-economic
486 components were tested with indirect structure testing (Barlas, 1996) and with respect to socio-economic
487 data on land use, crop choices and yields. Moreover, SDM had a coarse, spatially lumped representation
488 of the groundwater resources. For these reasons, the calibrated PBM described in Section 3.1 was
489 employed for the partial validation of the groundwater component of the system dynamics model
490 (Homer, 2012). In an experimental procedure, 42 runs were taken from the PBM to estimate the
491 parameters for the recharge and lateral flow rates in the system dynamics model. A detailed explanation
492 of the partial validation process is provided in Uygur (2023).

493 **3.4.2 Model output and analysis**

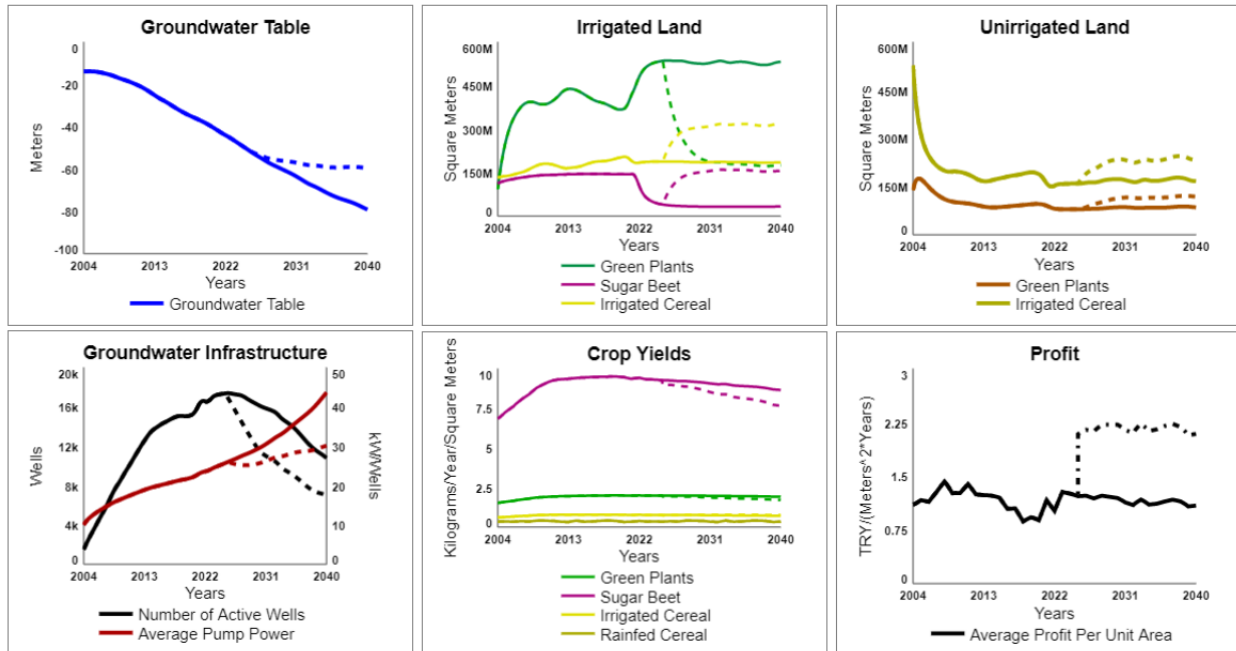
494 Figure 8 presents the results of the base run, compared to an integrated policy. In the historical period
495 (2004 through 2022), the unirrigated land shrinks as irrigation practices are adopted in the region. With
496 irrigation, crop yields increase over time, rendering irrigated agriculture more profitable and more
497 attractive. The increased demand for irrigation water requires infrastructure building; new wells are
498 drilled and the average pump power increases. As a result, increased extraction leads to a continuous
499 decrease in the groundwater table.

500 The policy scenario starting from 2025 includes interventions debated by various stakeholders and do not
501 necessarily reflect on potential enforcement problems. They are formulated as follows: the unit electricity
502 and green plants prices are decreased by 30% and 20% respectively, cereal and sugar beet prices are
503 increased by 60% and 75% respectively, a 2-year rotation scheme is enforced for the green plants (as the
504 current practice in some irrigation cooperatives in the region), 350 hm³ of surface water is supplied for

505 agricultural irrigation through the Mavi Tunnel, a recently constructed inter-basin water transfer project,
506 well drilling is prohibited in the 2025-2030 period, and a pumping quota of 90,000 m³/well/year is
507 enforced.

508 As can be viewed in Figure 8, the integrated policy scenario has rather counterintuitive results; it negates
509 the general belief in the basin that groundwater protection and maintaining the economic benefits of
510 agriculture (a win-win case for the environment and the farmers) are mutually exclusive. Under this
511 scenario, there is a significant improvement in groundwater level, compared to the base run. The crop
512 pattern changes drastically due to both the changes in the price and the crop rotation; while the area for
513 green plants decreases to one third of that of the base run, irrigated cereal and sugar beet become more
514 favorable for farmers with higher prices. The total irrigated land is reduced by 13% and converted to
515 unirrigated land. The decrease in overall water demand can be observed through the groundwater
516 infrastructure; both the number of active wells and the average pump power are lower under the policy
517 scenario compared to the base run. Additionally, even though lower irrigation levels lead to a declining
518 trend in sugar beet and green plants' yields, the average profit substantially increases under this policy
519 set, partly because of the increased crop prices and partly because the decrease in the unit electricity
520 price reduces farmers' overall costs.

521 Indeed, there might be numerous other policy scenarios that different stakeholders may consider superior
522 to this one, which implies that the "best" policy option is value-laden and dependent on the prioritized
523 goals. For this reason, the model interface was a useful tool at the third workshop, where a diverse
524 stakeholder group discussed their favored policies and their implications in the long run. Discussions of
525 issues that are 'elephant in the room', such as the number of unlicensed wells in the basin and crop
526 irrigation rates were facilitated through experimentation with the SDM by the end users and the water
527 managing authorities.



528

529 Figure 8: Key system variables in base run (solid lines) vs. the integrated policy scenario (dotted lines)

530

531 **4. Discussion**

532 The PBM, DDM and SDM approaches applied to the water-stressed Konya Closed Basin bring forth
 533 significant outcomes for sustainable groundwater management. Below it is argued that, while these
 534 modeling approaches have considerable differences with respect to their purposes, model formulations,
 535 model building processes, information sources, and validity and success criteria, their specific
 536 implementation in Konya has been interdependent and demonstrates potential complementary. In this
 537 section, first the generic differences and similarities of the three approaches are introduced. After that,
 538 their strengths and limitations, with reference to their specific implementation in Konya, are summarized.
 539 Last, the issues related to interdependence and complementarity during their implementation are
 540 summarized.

541 **4.1 Differences and similarities**

542 **4.1.1 Overarching purposes**

543 The PBM serves as an important tool that can aid in management of groundwater resources. It aims at
544 understanding the hydrogeologic conditions, constrained by available site-specific data, for the study area
545 that is highly dependent on groundwater resources. By considering alternative scenarios of exogenous
546 changes such as pumping rates, land use and climate, the PBM, after calibration against observed
547 historical data, can be utilized for predictive purposes, for example by estimating future groundwater
548 levels and water budget. The DDM aims for a practically applicable, fast operating tool for predictive
549 purposes. In that respect, it yields a useful tool to perform ‘what if’ analyses of the impact of exogenous
550 inputs (ex. pumping and precipitation) on resulting outputs (ex. groundwater heads). SDM aims at policy
551 exploration and design with respect to selected key performance indicators. It can internally generate
552 socio-economic scenarios (e.g., increasing pumping) under a selected set of exogenous choices (e.g., crop
553 prices). Though it is used to generate foresight to the future, it is rarely used for point-in-time predictive
554 purposes, fundamentally because its behavior is generated by its high order, multi-loop nonlinear
555 structure rather than exogenous drivers of change.

556 **4.1.2 Mathematical formulations and computational effort**

557 PBM is built on contemporary theory in the mature field of hydrogeology with decades of experimental
558 work and analyses. The model is formulated with spatially distributed partial differential equations, solved
559 numerically on computer platforms, often requiring long computation times. It is not a black box, but a
560 causal modeling approach relying on the latest representation of hydrogeologic conditions and flow
561 through porous media processes. DDM screens the regularity of patterns or intricate relationships among
562 huge data sets using machine learning techniques, statistical analyses, and pattern recognition algorithms,
563 amenable to fast computation. Differently from PBM, it operates by adopting a black-box approach that
564 prioritizes predictive accuracy over explicit causal or theoretical understanding among the variables of

565 concern. SDM builds on elicitation of variables with empirical/theoretical meaning and assessment of their
566 causal connection and feedback processes formulated in terms of differential equations. In non-spatial
567 SDM with aggregated agents, computation time is hardly a constraint. Individual mathematical
568 formulations are based on multiple theories, as well as controllable empirical truths. In this regard, SDM
569 is more similar to PBM with its causal/theoretical nature.

570 **4.1.3 Modeling process**

571 Available PBMs are often highly complex, spatially distributed constructs of the governing hydrogeological
572 processes. They are often packaged in re-usable codes to be implemented for different cases and for
573 relatively different purposes. Typical applications of PBM do not cover model building from scratch but
574 utilize available, already verified codes. For each application of DDM, tasks, such as data collection and
575 preprocessing (handling missing values, removing outliers, and normalization), are required. Feature
576 selection or engineering is then performed to extract the most informative variables for the modeling
577 task. An appropriate modeling technique is subsequently selected from a wide range of approaches such
578 as regression models and ANNs, tailored to the nature of the problem. The SDM is typically built from
579 scratch for a specific model purpose. In the case of SDM, model building benefits from generic model
580 structures and archetypes that have been developed for models with relatively different purposes,
581 theories from diverse disciplines of interest for the specific project, and empirical knowledge often
582 enhanced with various forms of stakeholder involvement.

583 **4.1.4 Data needs and use**

584 Respective differences in purposes, model formulation and modeling processes create different needs
585 and uses of data for each of the three approaches. Availability of large-scale quantitative data enhances
586 the usefulness and predictive capacity of PBM. On the other hand, because of its reliance on widely
587 studied processes, PBM is not totally bound to data availability. Based on its theoretical structure, it can

588 operate with limited direct measurements supplemented with inferred or measured data derived from
589 previous studies performed under similar hydrogeologic conditions, albeit with some increase in level of
590 uncertainty. Unlike traditional approaches that rely heavily on predefined mathematical equations or
591 theoretical frameworks, DDM derives its insights directly from the data itself. This means that the
592 effectiveness and reliability of DDM are entirely contingent upon the quality, quantity, and
593 representativeness of the data available. Every aspect of the modeling process, from model selection to
594 parameter tuning, hinges on the characteristics of the dataset at hand. Therefore, meticulous attention is
595 paid to data collection, preprocessing, and feature engineering to ensure that the data accurately reflects
596 the underlying phenomena being modeled. SDM benefits from a mix of quantified (numeric) and
597 qualitative data and often operates with limited data sets compared to the needs of its holistic, large
598 boundary view of socio-environmental systems. Whenever quantitative data sets are available, they are
599 primarily used to set the problem orientation and reference behavior of the model. While prior estimation
600 of model parameters is performed, due to limited data and structural uncertainty, many parameters are
601 indirectly estimated through model simulation tests.

602 **4.1.5 Validity and success criteria**

603 Validity and success criteria of the three modeling approaches are also differentiated. PBM is built on
604 established and developing theories, often relying on re-usable, therefore tested, computer codes. Hence,
605 for most PBM applications, the model structure and code is not rigorously scrutinized. The model quality
606 and usefulness is tested through model calibration whereby model input parameters characterized with
607 high levels of uncertainty are adjusted, within acceptable constraints, to improve agreement between
608 model output and key dependent variables subjected to imposed boundary and initial conditions. The
609 model calibration process benefits from sensitivity analyses performed to assess the behavior and
610 sensitivity of the model output to key model input parameters. The validity of a DDM is typically assessed
611 by testing its performance on unseen data, achieved through a process of training-validation split and

612 subsequent evaluation. Initially, the available data is partitioned into a training set, used to train the
613 model, and a validation set, used to evaluate its performance. The model is then trained on the training
614 set, learning patterns and relationships within the data. Following training, the model predictions are
615 compared to the actual values in the validation set using appropriate performance metrics. Additionally,
616 the model can be tested on entirely new, unseen data to validate its generalization ability. In SDM,
617 although the model is a causal representation of the system being analyzed, the underlying theories can
618 be insufficient, ambiguous, and the socio-environmental reality overly complex to be sufficiently
619 formulated in model structure. Therefore, every case specific application of SDM needs to go through the
620 process of confidence building on its theory-like structure. Hence, first structural validation tests are
621 implemented. Models with sufficient confidence in model structure then go through behavior validation
622 tests to check whether the model produced patterns match with real life data sets. Beyond visual,
623 descriptive judgements on pattern matching, statistical measures for means, trends, periods and
624 amplitudes are commonly used.

625 **4.2. Demonstrated strengths and limitations**

626 The versatility in incorporating various data types enhances the PBM's accuracy, adaptability, and
627 reliability, making it a valuable tool for simulating and predicting the complex dynamics of groundwater
628 systems in response to different scenarios and external factors. Its reliance on well-established physical
629 processes enhances user confidence and its reception among stakeholders and water resources
630 managers. The PBM developed for the whole Konya Closed Basin provided an estimate of the temporal
631 water balance for past conditions and future scenario analyses, to serve as one of crucial information for
632 sustainable groundwater management. It utilized geologic maps, well logs and pumping test data to
633 define the spatial distribution of the hydrogeological properties of the aquifer system. The model
634 incorporated extensive precipitation data from various monitoring stations for simulating spatially and
635 temporally dependent meteorological conditions and long-term historical groundwater data for model

636 validation and calibration. The basin lacked direct measurements of groundwater extraction data, like
637 many aquifers worldwide (Butler et al., 2023). For that reason, the model used an indirect estimate of
638 groundwater extraction rates, with calculations based on the documented spatial distribution of
639 cultivated areas of the various crops, their water demands and on irrigation efficiency scenarios.

640 A limitation encountered during the development of PBM of groundwater resources was the need for an
641 estimate of aquifer recharge. Because of lack of reliable estimates of the recharge, this parameter was
642 designated as a calibration parameter. To circumvent this requirement, the developed PBM for the Konya
643 Closed Basin was extended to incorporate vertical water flow in the vadose zone in addition to horizontal
644 groundwater flow. As such, the recharge in this model was an intermediate model output calculated from
645 the applied surface water and evapotranspiration rate instead of defining it as input to the groundwater
646 model.

647 The calibrated PBM was then used to assess the impact of climate change and the impact of different
648 pumping scenarios management changes on the sustainability of groundwater resources. A limitation of
649 PBM at this stage of analysis was its dependence on exogenous scenarios, either perpetuating past trends
650 or building wishful futures, disregarding the socio-hydrological response to changing groundwater
651 conditions in the long term.

652 Recent advancements in smart tools like surrogate DDMs have bolstered decision support systems
653 capabilities remarkably by reducing computational time, which is useful for making timely decisions in
654 dynamic environments. Surrogate DDMs expedite results without compromising accuracy, thus
655 promoting cost-effectiveness, and improving organizational performance. The surrogate DDM for Konya
656 Closed Basin offered easily interpretable results and is integrated with a user-friendly interface, making it
657 accessible to governmental authorities, farmers, NGOs, and many other stakeholders indirectly involved

658 in the groundwater debate, without the requirement for an in-depth understanding of the complex
659 physical dynamics.

660 Despite their benefits, the black-box nature of surrogate DDMs, such as ANNs, may lead to skepticism
661 about their reliability, invoking distrust among the end users of the tool. Emerging methods, such as
662 incorporating physical constraints during training, aim to enhance model interpretability (Raissi et al.,
663 2019; Secci et al., 2024). Furthermore, if high-quality and sufficient data are not available to adequately
664 describe various scenarios within an appropriate training range, the DDM cannot be effectively trained.
665 In fact, one limitation of DDM in this study was its entire reliance on the PBM, calibrated to historical data,
666 for the generation of a consistent range of data needed for training the ANN to simulate different
667 scenarios that exceed the historical precipitation and water demand datasets. Nevertheless, once trained,
668 the DDM can prove to be beneficial for Decision Support System and can also potentially serve as a
669 calibration model for the SDM, thereby significantly reducing the computational burden associated with
670 this process.

671 SDM is a robust methodology for the analysis of dynamically complex problems with an endogenous focus.
672 SDM allows for a multi-sectorial approach in groundwater management, incorporating various elements
673 of socio-hydrological complexity. The SDM for Konya Closed Basin had focused on the administrative unit
674 Çumra in and benefited from stakeholder knowledge and judgment at various stages of the study. It
675 created future water use scenarios considering multiple system wide decisions in feedback, such as crop
676 choices, factor use, infrastructure development and technological efficiency changes in response to
677 changing groundwater conditions. For example, SDM relaxed the best irrigation practice and full resource
678 accessibility assumptions subtle in scenarios perpetuating the past or building wishful futures. In that
679 respect, SDM partly covers the factors leading to the paradox of irrigation efficiency (Grafton et al. 2018).
680 The model interface for policy exploration and reduced simulation time enabled its use in a workshop
681 setting to facilitate communication and learning among the stakeholders.

682 The SDM had a simplified representation for groundwater hydrology, in a non-spatial “box model”
683 accumulating the lateral and vertical recharge minus pumping on an annual basis. This overly simplified
684 representation of groundwater hydrology risked SDM’s messages being highly generic, rather than
685 empirically grounded. To overcome this limitation, SDM benefited from PBM for partial validation of its
686 hydrological component. Specifically, the flow parameters for SDM were estimated with simulation tests
687 running the SDM and PBM in parallel under experimentally designed conditions of isolated and
688 comprehensive flows with extremely high and extremely low irrigation conditions. The PBM outputs of
689 these experiments were used to estimate SDM flow parameters.

690 **4.3 Interdependence and potential complementarity**

691 The Konya Closed Basin application and ensuing discussion demonstrates the different modeling
692 perspectives can be utilized to examine water-related issues and identify optimal management strategies.
693 The PBM rooted in scientific principles, provided a detailed understanding of underlying processes that
694 are crucial for accurate representation of complex systems. The DDM offered a valuable alternative for
695 tasks requiring rapid predictions and optimization. The SDM examined long-term groundwater
696 management from a socio-hydrological perspective. The output of PBM is used to train DDM, and for
697 partial validation of SDM. DDM once calibrated, can also be used for partial validation of SDM. SDM driven
698 scenarios can be used by PBM and DDM to predict future groundwater conditions. In future applications,
699 researchers can pursue enhanced coupling of these three approaches.

700 **5. Conclusions**

701 Numerical modeling has evolved into a vital tool for the management of groundwater resources. This
702 paper traces the development over the past decades of three modeling approaches: PBM, DDM and SDM
703 for the analysis of groundwater systems. We discuss these three modeling approaches in terms of their
704 objectives, model formalism and computation, modeling process, data requirements, validation,

705 computational effort and model use. To illustrate the strengths and limitations of these approaches, we
706 describe their application to the Konya Closed Basin, a major agricultural region of Turkey that faces
707 significant water stress due to over-exploitation of groundwater resources.

708 The three modeling approaches examined in this study have significant differences in various dimensions.
709 The application to the Konya plain illustrates their strengths and limitations, but also their
710 interdependence and suggested complementarity of these approaches. Depending on the specific
711 purposes of individual modeling projects, one or more of these approaches can be pursued. By combining
712 the strengths of PBM, DDM, and SDM, a synergistic effect can be achieved. This integrated approach not
713 only enhances predictive accuracy but also fosters a comprehensive grasp of the studied phenomena.
714 Moreover, the synergy enables researchers and practitioners to navigate the trade-off between
715 computational efficiency and model fidelity, ensuring a balanced and informed decision-making process.
716 Therefore, the collaboration between PBM, DDM and SDM presents a nuanced and effective approach
717 for advancing modeling methodologies across diverse applications. To make use of their strengths, future
718 model development efforts can focus on enhanced coupling of these modeling approaches. Recognizing
719 the strengths and limitations of each approach, a hybrid modeling strategy emerges as a powerful path
720 forward.

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